1 Effects of personal air pollution exposure on asthma symptoms, lung function and air	way
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51 Abstract

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#### 53 Background

There is evidence that air pollution increases the risk of asthma hospitalisations and healthcare
utilisation, but the effects on day-to-day asthma control are not fully understood.

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### 57 **Objective**

58 We undertook a prospective single-centre panel study to test the hypothesis that personal air 59 pollution exposure is associated with asthma symptoms, lung function and airway 60 inflammation.

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## 62 Methods

Thirty-two patients with a clinical diagnosis of asthma were provided with a personal air 63 pollution monitor (Cairclip NO<sub>2</sub>/O<sub>3</sub>) which was kept on or around their person throughout the 64 65 12-week follow-up period. Ambient levels of NO<sub>2</sub> and particulate matter were modelled based upon satellite imaging data. Directly measured ozone, NO<sub>2</sub> and particulate matter levels were 66 obtained from a monitoring station in central Leicester. Participants made daily electronic 67 records of asthma symptoms, peak expiratory flow, and exhaled nitric oxide. Spirometry and 68 69 asthma symptom questionnaires were completed at fortnightly study visits. Data were analysed 70 using linear mixed effects models and cross-correlation.

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#### 72 **Results**

Cairclip exposure data were of good quality with clear evidence of diurnal variability and a
missing data rate of approximately 20%. We were unable to detect consistent relationships
between personal air pollution exposure and clinical outcomes in the group as a whole. In an

resploratory subgroup analysis, total oxidant exposure was associated with increased daytimesymptoms in women but not men.

## 79 Conclusions and clinical relevance

We did not find compelling evidence that air pollution exposure impacts on day-to-day clinical control in an unselected asthma population, but further studies are required in larger populations with higher exposure levels. Women may be more susceptible than men to the effects of air pollution, an observation which requires confirmation in future studies.

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101 Introduction

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Ambient air pollution exposure contributes substantially to the global burden of disease, with 103 104 particulate matter pollution being the fifth-ranking mortality risk factor worldwide in 2015 [1]. The majority of this excess mortality is attributable to cardiovascular and respiratory conditions 105 106 such as ischaemic heart disease and chronic obstructive airways disease. Air pollution is mainly 107 derived from the combustion of fossil fuels and comprises a number of distinct chemical constituents. These include gaseous pollutants such as nitrogen oxides and ozone, and 108 109 particulate matter which may be further sub-divided into fine (aerodynamic diameter  $< 2.5 \mu m$ ,  $PM_{2.5}$ ) and coarse (aerodynamic diameter between 2.5 and 10µm,  $PM_{10}$ ). 110

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There is at present limited information available regarding the effects of air pollution exposure on individual patients in a real-life setting. This is primarily due to the difficulty of accurately evaluating personal exposures and correlating these with day-to-day variations in disease control. However, increasing miniaturisation of air quality monitors over the past five years has now made personal exposure monitoring a realistic prospect [2]. Furthermore, satellite remote sensing with geospatial modelling of ground-level air pollution is emerging as an important source of exposure data in epidemiological and cohort studies [3].

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Evidence from epidemiological and experimental studies suggests that air pollution may contribute to exacerbations of asthma through oxidative injury to the airways, increased sensitisation to aeroallergens, airway remodelling and airway hyperresponsiveness [4]. We therefore hypothesised that personal air pollution exposure is associated with day-to-day variations in asthma symptoms, lung function and airway inflammation. In order to test this hypothesis we undertook a prospective single-centre panel study, making use of portablesensors and satellite-based geospatial modelling.

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### 129 Methods

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#### 131 *Participants*

Thirty-two patients with asthma were recruited from general respiratory and specialist clinics 132 133 at Glenfield Hospital, Leicester. Patients were seen at baseline in the stable state, with no changes having been made to their regular inhaled or oral asthma therapy within the preceding 134 six weeks. All participants were never smokers or ex-smokers with less than 10 pack years' 135 136 smoking history. Asthma was diagnosed in a secondary or tertiary care setting according to British Thoracic Society guidelines [5]. The study was approved by the National Research and 137 Ethics Committee – East Midlands, Leicester, and all participants gave their written informed 138 consent. Participants were recruited and studied on a rolling basis from May 2016 to April 139 2017. 140

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## 142 *Exposure measurements*

Participants were provided with a Cairclip  $NO_2/O_3$  monitor (Cairpol, La Roche-Blanche, France), a cylindrical unit with a weight of 55g, length of 62mm and diameter of 32mm. The unit uses an electrochemical sensor to sample total oxidants (the sum of  $NO_2$  and  $O_3$ ) in the ambient air once per minute, and stores readings with an accurate date and time stamp. Cairclip  $NO_2/O_3$  has undergone detailed technical validation by the Joint Research Centre of the European Commission [6] and the United States Environment Protection Agency [7]. The Cairclip unit was kept within a fabric holder which could be easily attached to a belt or handbag. Participants were instructed to keep the Cairclip on their person when outdoors, and in close proximity to them when indoors. The battery life of the unit was approximately 24 hours and it therefore required charging overnight using a universal serial bus (USB) cable. The same USB connection was used to extract data from the unit at study visits. The unit was set to store 154 15 minute average readings so as not to exceed its data storage capacity between the twoweekly study visits.

Directly measured daily average levels of NO<sub>2</sub>, O<sub>3</sub> and PM<sub>2.5</sub> at the primary air quality 157 158 monitoring station in central Leicester were obtained from the Department for the Environment, Food and Rural Affairs (DEFRA) (https://uk-air.defra.gov.uk/data/). Modelled 159 average daily levels of NO<sub>2</sub>, PM<sub>10</sub> and PM<sub>2.5</sub> were produced by the Earth Observation Science 160 Group at the University of Leicester with a spatial resolution of 1 km<sup>2</sup>. This dataset was derived 161 from satellite data from the European ENSEMBLE model, which corresponds to the median 162 of seven European air quality models provided by Copernicus Atmosphere [8, 9]. The 163 164 ENSEMBLE model provides surface levels of daily mean values of NO<sub>2</sub>, PM<sub>2.5</sub> and PM<sub>10</sub> at a spatial resolution of 0.1° (around 7 km). These were interpolated onto a 1 km grid scale based 165 upon DEFRA national pollution climate mapping [10]. Figure 1 shows annual mean NO<sub>2</sub> 166 concentrations in 2016 for the City of Leicester calculated using this methodology. Exposures 167 for each participant were calculated based upon the grid square in which their home address 168 169 was located.

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171 Daily home monitoring

In order to provide daily time series of clinical outcomes, participants were provided with an
electronic symptom diary and peak expiratory flow (PEF) meter (Micro Diary; Carefusion,
Basingstoke, UK) and a portable exhaled nitric oxide (FeNO) monitor (NOBreath; Bedfont,

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175 Maidstone, UK). Participants used the Micro Diary unit twice per day in the morning and evening to record visual analogue scores for cough, breathlessness and wheeze over the 176 preceding overnight or daytime period. These were recorded on an integer scale from zero (no 177 symptoms) to ten (severe symptoms). Total night-time and daytime symptom scores (on an 178 integer scale of zero to thirty) were calculated as the sum of the three visual analogue scores 179 for the preceding overnight or daytime period. Participants also undertook three PEF 180 measurements in the morning and evening using the same unit, and the best of the three 181 readings was recorded on each occasion. 182

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184 Two measurements of FeNO were made each morning using the NOBreath unit, at an 185 exhalation rate of 50 ml/s. These were manually entered by patients into the Micro Diary since 186 NOBreath does not have the facility for storing multiple test results. The average of the two 187 daily readings was utilised in statistical analysis. We have previously shown that home FeNO 188 monitoring using NOBreath produces reproducible results which correlate well with blood and 189 sputum eosinophil counts [11].

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191 *Study visit schedule* 

Participants attended a baseline visit at which clinical and demographic details were recorded and the home monitoring equipment was introduced and explained. Participants were then followed up for a total of 12 weeks and attended study visits at 2-weekly intervals, at which time data were downloaded from the Cairclip and Micro Diary units. Participants also completed the Asthma Control Questionnaire [12] and Asthma Quality of Life Questionnaire [13], and underwent spirometry. Spirometry was performed according to American Thoracic Society / European Respiratory Society guidelines [14] with predicted values calculated using Global Lung Initiative reference equations [15]. Short and long-acting bronchodilators werewithheld for 4 and 12 hours respectively before study visits.

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#### 202 Statistical analysis

203 Statistical analysis was performed using SPSS 24 (IBM, Armonk, New York, USA). 204 Relationships between exposure and outcome variables were analysed using linear mixed 205 effects models and cross-correlation, with the two techniques giving complementary 206 information.

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Linear mixed effects models were used to assess the relationship between linear trends in the 208 209 exposure and outcome variables over the study period at a group level, with separate models 210 constructed for each clinical outcome. Clinical outcomes were entered as dependent variables 211 in the model, with exposures specified as fixed effects. An autoregressive covariance structure was specified to account for the observed autocorrelation in the clinical time series. For 212 outcomes measured daily (symptom scores, PEF and FeNO) the corresponding exposure 213 variable was the average total oxidant reading on the same day measured using the Cairclip 214 unit. For outcomes measured at study visits the average total oxidant over the previous two 215 weeks was used as the corresponding exposure variable for spirometry and the Asthma Quality 216 of Life Questionnaire, and over the previous week for the Asthma Control Questionnaire. 217

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Cross-correlation was used to assess the relationship between day-to-day variations in exposure and outcome variables in individual participants over a range of lag times (-7 to +7 days).
Cross-correlation results were assessed individually, and were combined across the group to give the mean and 95% confidence interval cross-correlation at each lag time. Exploratory

subgroup analyses were performed, stratifying the group by sex and smoking history (ex-smokers versus never smokers).

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## 226 Statistical power calculation

The primary analysis was estimation of the effect of average total oxidant exposure on peak 227 expiratory flow across the cohort, analysed using a linear mixed effects model. Calculation of 228 statistical power was performed by Dr. Matthew Richardson (University of Leicester) using a 229 simulation approach, based upon an internal pilot comprising the first six participants in the 230 231 study. A linear mixed effects model was fitted to the pilot data; the dependent variable was morning PEF (PEFam) and the independent variable was the daily average personal total 232 oxidant concentration (AverageOX). A random intercept for each subject was also included in 233 234 the model. Estimates for the fixed effect intercept, variance of the random intercept and 235 variance of the error term were obtained from the pilot sample. Values for PEFam were then generated according to the regression equation derived from the pilot data, but allowing the 236 237 regression coefficient for AverageOX to vary. The random intercept for each subject was assumed to follow a normal distribution, and a normally distributed error term was included in 238 the regression equation. The p-value for the regression coefficient of AverageOX was obtained 239 using 100 simulations, with sample size fixed at n=30. Based on the simulation results, it was 240 estimated that this sample size would be sufficient to detect an effect size of AverageOx (ppb) 241 242 on PEFam (L/min) of approximately 0.7 units with power=80% at the 5% significance level.

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- 244
- 245 **<u>Results</u>**

247 Thirty-one participants completed the study, with one participant withdrawing at an early stage due to technical problems with the Cairclip unit. Clinical and demographic characteristics of 248 the participants are shown in Table 1. Participants were followed up for an average of 78 days, 249 250 with a total of 2406 person-days of follow-up. The average missing data rate for daily monitoring was 15.5% for morning readings, 21.7% for evening readings and 20.6% for 251 252 Cairclip total oxidant data. Missing Cairclip data were usually caused by complete discharge of the internal battery, due to the unit being left off charge overnight. This resulted in loss of 253 254 the timestamp and necessitated return of the unit to the study centre to be reset. Figure 2 shows 255 an example time series of personal total oxidant exposure in a single participant measured using Cairclip. This shows the typical diurnal pattern of exposure, with peak levels during daylight 256 257 hours, and considerable day-to-day variability with regard to average daily exposure. Figure 3 258 shows the average total oxidant levels measured using Cairclip at different times of the day 259 over 1911 person-days of observations. A peak is seen during day-time hours, with additional small peaks corresponding to the morning and evening rush hours on weekdays. Although 260 261 exposure levels were slightly lower on weekends compared to weekdays, clinical outcomes did not differ significantly across the seven days of the week, as shown in Supplementary Table 262 S1. 263

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Table 2 shows the results of linear mixed effects models in which the fixed effects of average daily total oxidant exposure measured using Cairclip on clinical outcomes were estimated over the whole study group. In each case, no clinically important or statistically significant effects were observed. Cross-correlation analysis did not reveal any consistent associations between exposure and outcome variables in the study group as a whole, as shown in Figure 4 and Supplementary Figures S1-S6. When the group was stratified by sex, we observed a positive correlation on the same day (lag 0) between total oxidant exposure using Cairclip and daytime symptoms in females, but not in males (Figure 5). Stratifying the group by smoking history didnot reveal any significant associations.

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276 Discussion

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We report the first panel study of the effects of personal air pollution exposure on adults with 278 asthma in which exposures and outcomes were assessed concurrently over an extended 12-279 280 week period. We assessed air pollution exposure using three separate methodologies, namely (i) a small portable sensor carried on the person (Cairclip NO<sub>2</sub>/O<sub>3</sub>), (ii) modelled data based on 281 satellite imaging, and (iii) directly measured ambient levels at a central monitoring station. We 282 283 measured a comprehensive panel of daily clinical outcomes, including asthma symptom scores, 284 peak expiratory flow and exhaled nitric oxide. Most previous longitudinal studies of air pollution effects on asthma have relied solely on central monitoring stations to estimate the 285 exposure of an entire community. A small number of studies have utilised personal air pollution 286 sensors [16-19], but due to the bulky nature of the monitoring equipment the follow-up duration 287 has necessarily been limited. We have for the first time demonstrated the feasibility of long-288 term personal air pollution exposure monitoring. 289

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We did not observe strong associations between exposures and clinical outcomes, although an exploratory sub-group analysis suggested the possibility of increased daytime symptoms in women exposed to air pollution. A number of possible explanations may be suggested to explain our findings. The participants in our study were recruited from Leicester and the surrounding area. In order to increase the generalisability of our results we did not restrict participation to individuals residing in heavily polluted districts. Therefore, a number of study

297 participants lived in rural areas and may not have had sufficient air pollution exposure to impact on their asthma control. Indeed, average peak daily exposure to total oxidant measured using 298 Cairclip monitors was approximately 10 ppb in our study cohort, which is well within UK and 299 300 European air quality limits [20]. Future studies should therefore focus more specifically on high-risk populations, such as those living and working in inner-city areas. Moreover, our data 301 collection and statistical methods relied on detecting associations between exposure and 302 outcome variables produced by day-to-day variability and linear trends arising naturally. We 303 did not impose an experimental change in exposure, as for instance was done by McCreanor et 304 305 al, who measured changes in forced expiratory volume in one second caused by walking down a busy London street (Oxford Street) compared to a nearby park (Hyde Park) [21]. Therefore, 306 307 the lack of observed associations could also have been due to low levels of day-to-day 308 variability in exposures and/or clinical outcomes.

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With 31 participants we had sufficient statistical power to detect a 21 L/min change in peak expiratory flow in association with a 30 ppb change in total oxidant exposure. It is possible that a true effect existed but that it was too small to detect with our sample size. Indeed, previous cohort studies have often reported smaller effect sizes than this [22]. However, it has been shown that changes in peak expiratory flow of less than 20 L/min are unlikely to result in perceptibly increased symptoms [23]. Therefore we considered that our sample size was sufficient to detect clinically important effects on lung function.

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We endeavoured to reduce the missing data rate by seeing participants frequently at the study centre (every 2 weeks) and providing them with contact details of investigators who could quickly resolve technical issues with the home monitoring equipment. Nevertheless,

approximately 15-20% of daily exposure or outcome data were missing, and this may havereduced the chance of observing positive associations.

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324 In this study we particularly investigated the effects or air pollution exposure on day-to-day variations in asthma symptoms, lung function and airway inflammation. It is possible that the 325 major effects of air pollution on asthma relate to other aspects of the condition which our study 326 was not designed to measure, such as exacerbations, hospital admissions and lung function 327 decline. Indeed, a recent meta-analysis of 87 epidemiological studies conducted in multiple 328 329 cities worldwide concluded that short-term exposure to a number of air pollutants results in significantly increased risks of asthma-related hospital admissions [24]. It has been reported in 330 a recent large population-based cohort study that cumulative air pollution exposure is a 331 332 significant risk factor for progression from asthma to chronic obstructive pulmonary disease [25]. While significant at a population level, these effects may be too subtle to detect in small 333 panel studies. 334

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Susceptibility to the effects of air pollution varies between individuals and is most likely related 336 to genetic factors [26]. Therefore, it is possible that clinically significant effects of air pollution 337 may be confined to a small but important group of individuals with a combination of biological 338 susceptibility and higher-than-average levels of exposure. Future research should seek to detect 339 340 individuals at particular risk from air pollution and to develop strategies to mitigate this risk. Our data suggest that women may be more susceptible to the effects of air pollution than men, 341 an observation which is shared with a number of previous studies [27], and which requires 342 343 further investigation.

We chose to use total oxidant levels measured using the Cairclip  $NO_2/O_3$  monitor as our 345 primary exposure variable. This device was chosen due to a combination of portability, 346 relatively low cost, and good measurement properties (precision, linearity, limits of detection) 347 under laboratory and field conditions [6, 7, 28]. However, these validation studies were 348 performed with the sensor in the static state, and it is known that movement of air pollution 349 sensors can have a significant effect on the measurements [29]. A further consideration for 350 mobile monitoring is the sensor response time, which although relatively fast at between 1.5 351 and 4 minutes for Cairclip [6, 7] may still not be sufficient to detect very transient increases in 352 353 air pollution exposure. Therefore sensors used in future personal exposure studies should first undergo validation under realistic field conditions [30]. Jaio et al found that Cairclip NO<sub>2</sub>/O<sub>3</sub> 354 accurately measured the sum of NO<sub>2</sub> and O<sub>3</sub> (total oxidants) when compared against reference 355 356 sensors [31]. However, when an estimate of NO<sub>2</sub> was derived by subtracting reference values 357 of O<sub>3</sub> from Cairclip total oxidant measurements, these did not correlate with reference values of NO<sub>2</sub>. This suggests that Cairclip NO<sub>2</sub>/O<sub>3</sub> measurements are not entirely additive and cannot 358 be used to reliably measure the individual NO<sub>2</sub> and O<sub>3</sub> components. This is a potential 359 limitation of the study, since it is possible that measuring NO<sub>2</sub> or ozone alone would have 360 resulted in greater correlation with clinical outcomes. However, Cairclip NO<sub>2</sub> was still in 361 development when our study commenced and is less well validated than Cairclip NO<sub>2</sub>/O<sub>3</sub> [28]. 362 There are now a number of sensors in development or commercially available, measuring either 363 364 particulate matter [32] or gaseous pollutants [33, 34], which may be used in future personal exposure studies. It is likely that the measurement properties of these sensors will continue to 365 improve as technology develops. 366

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368 The current challenges and opportunities in the use of sensors to estimate environmental 369 exposures have been recently summarised by Loh *et al* [2]. While small portable air pollution 370 monitors have not yet been fully validated in the ambulatory setting, an alternative approach is 371 to collect location and activity data using Global Positioning System-enabled smartphone 372 applications, and to link these to ambient levels of air pollution measured using static ground-373 level monitors or satellite imaging. Furthermore, personal exposure to air pollution is 374 dependent not only on ambient levels but also on the activity being undertaken and the rate of 375 ventilation, which can also be estimated using wearable technology.

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In conclusion, we were unable to find compelling evidence that air pollution exposure impacts 377 378 on day-to-day clinical control in an unselected asthma population, but these results should be taken in the context of the relatively small study population and the low levels of exposure 379 experienced. Future studies should focus on selected individuals who are highly exposed to air 380 381 pollution due to their lifestyle or place of residence. Our results suggest the possibility of 382 increased susceptibility to air pollution in women compared to men, but this observation requires confirmation in further prospective studies. We have for the first time demonstrated 383 384 the feasibility of long-term personal exposure monitoring. Sensor technology and geospatial data processing is developing rapidly and will in the near future enable us to further understand 385 the relationship between health and the environment in a variety of long-term medical 386 conditions. 387

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# 395 **<u>References</u>**

397	1)	Cohen AJ, Brauer M, Burnett R, Anderson HR, Frostad J, Estep K, Balakrishnan K,
398		Brunekreef B, Dandona L, Dandona R, Feigin V, Freedman G, Hubbell B, Jobling A,
399		Kan H, Knibbs L, Liu Y, Martin R, Morawska L, Pope CA 3rd, Shin H, Straif K,
400		Shaddick G, Thomas M, van Dingenen R, van Donkelaar A, Vos T, Murray CJL,
401		Forouzanfar MH. Estimates and 25-year trends of the global burden of disease
402		attributable to ambient air pollution: an analysis of data from the Global Burden of
403		Diseases Study 2015. Lancet. 2017; 389(10082): 1907-1918.
404		
405	2)	Loh M, Sarigiannis D, Gotti A, Karakitsios S, Pronk A, Kuijpers E, Annesi-Maesano I,
406		Baiz N, Madureira J, Oliveira Fernandes E, Jerrett M, Cherrie JW. How sensors might
407		help define the external exposome. Int J Environ Res Public Health. 2017; 14(4): 434.
408		
409	3)	Sorek-Hamer M, Just AC, Kloog I. Satellite remote sensing in epidemiological studies.
410		Curr Opin Pediatr. 2016; 28(2): 228-34.
411		
412	4)	Guarnieri M, Balmes JR. Outdoor air pollution and asthma. Lancet. 2014; 383(9928):
413		1581-92.
414		
415	5)	British Thoracic Society/Scottish Intercollegiate Guidelines Network, 2016. British
416		Guideline on the Management of Asthma. Available from: https://www.brit-
417		thoracic.org.uk/standards-of-care/guidelines/btssign-british-guideline-on-the-
418		management-of-asthma/
419		

420	6)	Joint Research Centre, 2013. Report of laboratory and in of micro-sensor for monitoring					
421		ambient air pollution: CairClipO3/NO2 of CAIRPOL. Available from:					
422		http://publications.jrc.ec.europa.eu/repository/					
423							
424	7)	Environmental Protection Agency: Washington, DC, USA, 2014. Sensor Evaluation					
425		Report; Report EPA/600/R-14/143. Available from: https://cfpub.epa.gov/si/					
426							
427	8)	Peuch VH, Engelen R, Simmons A, Lahoz W, Laj P, Galmarini S. Monitoring					
428		atmospheric composition and climate, research in support of the Copernicus/GMES					
429		atmospheric service, Special Issue. Atmos. Chem. Phys. 2014.					
430							
431	9)	Copernicus Atmosphere, 2016. Final Report MACC-III (Monitoring Atmospheric					
432		Composition and Climate 3). Available from: <u>http://atmosphere.copernicus.eu/reports</u>					
433							
434	10)	Department for the Environment, Food and Rural Affairs, 2015. 1 km DEFRA national					
435		pollution climate mapping for NO <sub>2</sub> , PM <sub>10</sub> and PM <sub>2.5</sub> . Available from: <u>https://uk-</u>					
436		air.defra.gov.uk/data/pcm-data					
437							
438	11)	Nanda CR, Singapuri A, Soares M, Monteiro W, Siddiqui S, Gonem S. Domiciliary					
439		exhaled nitric oxide and eosinophilic airway inflammation in adults with asthma. Eur					
440		Respir J. 2016; 48(1): 242-4.					
441							
442	12)	Juniper EF, O'Byrne PM, Guyatt GH, Ferrie PJ, King DR. Development and validation					
443		of a questionnaire to measure asthma control. Eur Respir J. 1999; 14(4): 902-7.					
444							

445	13)	Juniper EF, Buist AS, Cox FM, Ferrie PJ, King DR. Validation of a standardized
446		version of the Asthma Quality of Life Questionnaire. Chest. 1999; 115(5): 1265-70.
447		
448	14)	Miller MR, Hankinson J, Brusasco V, Burgos F, Casaburi R, Coates A, Crapo R,
449		Enright P, van der Grinten CP, Gustafsson P, Jensen R, Johnson DC, MacIntyre N,
450		McKay R, Navajas D, Pedersen OF, Pellegrino R, Viegi G, Wanger J; ATS/ERS Task
451		Force. Standardisation of spirometry. Eur Respir J. 2005; 26(2): 319-338.
452		
453	15)	Quanjer PH, Stanojevic S, Cole TJ, Baur X, Hall GL, Enright PL, Hankinson JL, Ip
454		MS, Zheng J, Stocks J. on behalf of the ERS Global Lung Function Initiative. (2012)
455		Multi-ethnic reference values for spirometry for the 3-95 year age range: the global
456		lung function 2012 equations. Eur Respir J. 2012; 40(6): 1324-1343.
457		
458	16)	Delfino RJ, Staimer N, Gillen D, Tjoa T, Sioutas C, Fung K, George SC, Kleinman
459		MT. Personal and ambient air pollution is associated with increased exhaled nitric oxide
460		in children with asthma. Environ Health Perspect. 2006; 114(11): 1736-43.
461		
462	17)	Delfino RJ, Staimer N, Tjoa T, Gillen D, Kleinman MT, Sioutas C, Cooper D. Personal
463		and ambient air pollution exposures and lung function decrements in children with
464		asthma. Environ Health Perspect. 2008; 116(4): 550-8.
465		
466	18)	Spira-Cohen A, Chen LC, Kendall M, Lall R, Thurston GD. Personal exposures to
467		traffic-related air pollution and acute respiratory health among Bronx schoolchildren
468		with asthma. Environ Health Perspect. 2011; 119(4): 559-65.
469		

470	19)	Smargiassi A, Goldberg MS, Wheeler AJ, Plante C, Valois MF, Mallach G, Kauri LM5,				
471		Shutt R, Bartlett S, Raphoz M, Liu L. Associations between personal exposure to air				
472		pollutants and lung function tests and cardiovascular indices among children with				
473		asthma living near an industrial complex and petroleum refineries. Environ Res. 2014;				
474		132: 38-45.				
475						
476	20)	Department for the Environment, Food and Rural Affairs, 2017. National air quality				
477		objectives. Available from: https://uk-air.defra.gov.uk/air-pollution/uk-eu-limits				
478						
479	21)	McCreanor J, Cullinan P, Nieuwenhuijsen MJ, Stewart-Evans J, Malliarou E, Jarup L,				
480		Harrington R, Svartengren M, Han IK, Ohman-Strickland P, Chung KF, Zhang J.				
481		Respiratory effects of exposure to diesel traffic in persons with asthma. N Engl J Med.				
482		2007; 357(23): 2348-58.				
483						
484	22)	United States Environmental Protection Agency, 2013. Integrated Science Assessment				
485		for Ozone and Related Photochemical Oxidants. Available from:				
486		https://www.epa.gov/isa/integrated-science-assessment-isa-ozone-and-related-				
487		photochemical-oxidants				
488						
489	23)	Santanello NC, Zhang J, Seidenberg B, Reiss TF, Barber BL. What are minimal				
490		important changes for asthma measures in a clinical trial? Eur Respir J. 1999; 14(1):				
491		23-7.				
492						
493	24)	Zheng XY, Ding H, Jiang LN, Chen SW, Zheng JP, Qiu M, Zhou YX, Chen Q, Guan				
494		WJ. Association between Air Pollutants and Asthma Emergency Room Visits and				

495 Hospital Admissions in Time Series Studies: A Systematic Review and Meta-Analysis. 496 PLoS One. 2015; 10(9): e0138146. 497 498 25) To T, Zhu J, Larsen K, Simatovic J, Feldman L, Ryckman K, Gershon A, Lougheed MD, Licskai C, Chen H, Villeneuve PJ, Crighton E, Su Y, Sadatsafavi M, Williams D, 499 500 Carlsten C; Canadian Respiratory Research Network. Progression from Asthma to Chronic Obstructive Pulmonary Disease. Is Air Pollution a Risk Factor? Am J Respir 501 Crit Care Med. 2016; 194(4): 429-38. 502 503 Rava M, Smit LA, Nadif R. Gene-environment interactions in the study of asthma in 26) 504 505 the postgenomewide association studies era. Curr Opin Allergy Clin Immunol. 2015; 506 15(1): 70-8. 507 27) Clougherty JE. A growing role for gender analysis in air pollution epidemiology. 508 509 Environ Health Perspect. 2010; 118(2): 167-76. 510 Duvall RM, Long RW, Beaver MR, Kronmiller KG, Wheeler ML, Szykman JJ. 511 28) Performance Evaluation and Community Application of Low-Cost Sensors for Ozone 512 and Nitrogen Dioxide. Sensors (Basel). 2016; 16(10). pii: E1698. 513 514 29) Lerner U, Yacobi T, Levy I, Moltchanov SA, Cole-Hunter T, Fishbain B. The effect of 515 ego-motion on environmental monitoring. Sci Total Environ. 2015; 533: 8-16. 516 517 Jerrett M, Donaire-Gonzalez D, Popoola O, Jones R, Cohen RC, Almanza E, de Nazelle 30) 518 A, Mead I, Carrasco-Turigas G, Cole-Hunter T, Triguero-Mas M, Seto E, 519

520		Nieuwenhuijsen M. Validating novel air pollution sensors to improve exposure
521		estimates for epidemiological analyses and citizen science. Environ Res. 2017; 158:
522		286-294.
523		
524	31)	Jiao W, Hagler G, Williams R, Sharpe R, Brown R, Garver D, Judge R, Caudill M,
525		Rickard J, Davis M, Weinstock L, Zimmer-Dauphinee S, Buckley K. Community Air
526		Sensor Network (CAIRSENSE) project: evaluation of low-cost sensor performance in
527		a suburban environment in the southeastern United States. Atmos Meas Tech. 2016; 9:
528		5281–5292.
529		
530	32)	Koehler KA, Peters TM. New Methods for Personal Exposure Monitoring for Airborne
531		Particles. Curr Environ Health Rep. 2015; 2(4): 399-411.
532		
533	33)	McKercher GR, Salmond JA, Vanos JK. Characteristics and applications of small,
534		portable gaseous air pollution monitors. Environ Pollut. 2017; 223: 102-110.
535		
536	34)	Peterson PJD, Aujla A, Grant KH, Brundle AG, Thompson MR, Vande Hey J, Leigh
537		RJ. Practical Use of Metal Oxide Semiconductor Gas Sensors for Measuring Nitrogen
538		Dioxide and Ozone in Urban Environments. Sensors (Basel). 2017; 17(7). pii: E1653.
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545 Table 1: Clinical and demographic characteristics

Age (years)	55.2 (13.1)
Sex	Female n=15 (48.4%) Male n=16 (51.6%)
Body mass index (kg/m <sup>2</sup> )	30.4 (6.7)
Positive atopic status* (n [%])	24 (77.4)
Duration of asthma (years)	32.2 (21.5)
Age of onset of asthma (years)	23.0 (21.4)
Number of asthma exacerbations in the previous year	2.4 (4.5)
Regular inhaled corticosteroid use (n [%])	29 (93.5)
Inhaled corticosteroid dose (beclometasone dipropionate equivalent [µg])	1000 (400-1600)
Regular oral prednisolone use (n [%])	7 (22.6)
Leukotriene receptor antagonist use (n [%])	7 (22.6)
Oral theophylline use (n [%])	3 (9.7)
Long-acting $\beta_2$ agonist use (n [%])	21 (67.7)
Asthma Control Questionnaire score	1.41 (1.15)
Asthma Quality of Life Questionnaire score	5.32 (1.19)
Forced expiratory volume in one second (% predicted)	78.6 (22.5)
Forced vital capacity (% predicted)	91.1 (17.3)
Ratio of forced expiratory volume in one second to forced vital capacity (%)	67 (12)
Exhaled nitric oxide at a flow rate of 50ml/s (ppb)	28.1 (22.3)

547 Data are n (%), mean (standard deviation) or median (interquartile range) unless specified otherwise.

\*Defined as a positive skin test for any of a panel of specified aeroallergens (grass pollen, tree pollen

549 [alder, silver birch, hazel], moulds [Aspergillus fumigatus, Alternaria tenius, Cladosporium,

550 *Penicillium notatum*], cat fur, dog dander, and house dust mite [*Dermatophagoides pteronyssimus*]).

# 552 **Table 2: Linear mixed models estimating fixed effects of total oxidant exposure on daily**

553 clinical outcomes

Outcome measure	Estimate of fixed effects*	95% Confidence Interval		P value
(units/range)		Lower bound	Upper bound	
Morning PEF (L/min)	0.038	-0.416	0.491	0.870
Evening PEF (L/min)	-0.088	-0.576	0.400	0.723
FeNO (ppb)	0.012	-0.138	0.162	0.875
Night-time symptoms (0-30)	0.0036	-0.0182	0.0253	0.746
Daytime symptoms (0-30)	0.0022	-0.0246	0.0290	0.872
ACQ (0-6)	0.0032	-0.0250	0.0314	0.822
AQLQ (1-7)	-0.0246	-0.0613	0.0121	0.187
Pre-bronchodilator FEV <sub>1</sub> (% predicted)	-0.213	-0.766	0.340	0.447
Post-bronchodilator FEV <sub>1</sub> (% predicted)	-0.140	-0.586	0.306	0.535
Pre-bronchodilator FVC (% predicted)	0.261	-0.316	0.839	0.372
Post-bronchodilator FVC (% predicted)	0.083	-0.590	0.757	0.807
Pre-bronchodilator FEV1/FVC (% predicted)	-0.0017	-0.0054	0.0021	0.385
Post-bronchodilator FEV1/FVC (% predicted)	0.0008	-0.0023	0.0038	0.618

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555 PEF = peak expiratory flow rate; FeNO = exhaled nitric oxide measured at an expiratory flow rate of

556 50ml/s; ACQ = Asthma Control Questionnaire score; AQLQ = Asthma Quality of Life Questionnaire

557 score;  $FEV_1$  = forced expiratory volume in on second.

\*Exposure variable was measured in parts per billion using Cairclip NO<sub>2</sub>/O<sub>3</sub> unit.

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# 563 Figure 1: Annual mean NO<sub>2</sub> concentrations in 2016 for the City of Leicester

564 Raw satellite pixels from the ENSEMBLE model data (0.1 degrees resolution in latitude and

longitude) were interpolated on a 1 km grid using spatial information from the DEFRA PCM

566 database [10]. Map data are from Google Maps.





- 576 Figure 2: Personal total oxidant exposure measured using a Cairclip NO<sub>2</sub>/O<sub>3</sub> unit in an
- 577 individual with asthma





## 593 Figure 3: Diurnal variation in total oxidant levels measured using Cairclip

Average total oxidant levels are shown during the 24-hour daily cycle over 1911 person-days
of observations. Results are presented for all days, weekdays (Monday to Friday), and
weekends (Saturday and Sunday).



## **Figure 4: Group-level cross-correlations between total oxidant levels measured using Cairclip and clinical outcomes**

Group-level cross-correlations are shown between total oxidant levels measured using Cairclip and clinical outcomes. The group mean and 95%
confidence interval of the cross correlation is shown for each lag time (days). A lag of zero refers to effects occurring on the same day; positive
lags refer to changes in outcome preceded by changes in exposure; negative lags refer to changes in exposure preceded by changes in outcome.

612 PEF = peak expiratory flow; FeNO = fractional exhaled nitric oxide.



#### 614 Figure 5: Cross-correlation between total oxidant levels measured using Cairclip and daytime symptoms stratified by sex 615

Cross-correlations are shown between total oxidant levels measured using Cairclip and daytime 616 symptoms, in female and male participants (Panels A and B respectively). The group mean and 617 95% confidence interval of the cross correlation is shown for each lag time (days). A lag of 618 zero refers to effects occurring on the same day; positive lags refer to changes in outcome 619 preceded by changes in exposure; negative lags refer to changes in exposure preceded by 620 changes in outcome. 621

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Figure S1: Group-level cross-correlations between directly measured ambient nitrogen dioxide and clinical outcomes



Figure S2: Group-level cross-correlations between directly measured ambient ozone and clinical outcomes



Figure S3: Group-level cross-correlations between directly measured ambient PM<sub>2.5</sub> and clinical outcomes



Figure S4: Group-level cross-correlations between modelled ambient nitrogen dioxide and clinical outcomes



Figure S5: Group-level cross-correlations between modelled ambient PM<sub>10</sub> and clinical outcomes



Figure S6: Group-level cross-correlations between modelled ambient PM<sub>2.5</sub> and clinical outcomes